

1 **Ingested plastic and trace element concentrations in Short-tailed Shearwaters (*Ardenna***
2 ***tenuirostris*)**

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17 **ABSTRACT**

18 Pollution of marine environments is concerning for complex trophic systems. Two anthropogenic
19 stresses associated with marine pollution are the introduction of marine plastic and their associated
20 chemicals (e.g., trace elements) which, when ingested, may cause harm to wildlife. Here we explore
21 the relationship between plastic ingestion and trace element burden in the breast muscle of Short-
22 tailed Shearwaters (*Ardenna tenuirostris*). We found no relationship between the amount of plastic
23 ingested and trace element concentration in the birds' tissues. Though the mass and number of
24 plastic items ingested by birds during 1969 - 2017 did not change significantly, trace element
25 concentrations of some elements (Cu, Zn, As, Rb, Sr and Cd), appeared to have increased in birds
26 sampled in 2017 compared to limited data from prior studies. We encourage policy which considers
27 the data gleaned from this sentinel species to monitor the anthropogenic alteration of the marine
28 environment.

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30 *Keywords:* *Ardenna tenuirostris*; ecological indicator; marine debris; muttonbird; plastic pollution;
31 toxicology

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33 Marine and terrestrial environments face ever-growing pressures in the human-induced
34 Anthropocene epoch (Crutzen, 2002; Waters et al., 2016). In the case of the ocean environment,
35 which fosters life and drives global climate (Hoegh-Guldberg and Bruno, 2010), there is growing
36 concern following predictions of further biodiversity loss and increasing pollution levels
37 (MacArthur et al., 2016; Worm, 2016). Understanding and mitigating pressures on marine
38 ecosystems typically requires two pieces of important information: data which provide an indicator
39 of historical (baseline) conditions, and follow-up (monitoring) data to identify any changes over
40 time and guide future research (Pauly, 1995). To achieve this in the vast, dynamic marine
41 environment, we can make use of sentinel (or indicator) species which can provide large volumes of
42 data at relatively low cost while being specific to an area of interest (Caro, 2010; Durant et al.,
43 2009; Van Franeker et al., 2011). As apex predators that faithfully return to breeding islands,
44 seabirds are an ideal sentinel for the marine environment, reflecting changes in prey abundance and
45 pollution, including a suite of chemicals (Mallory et al., 2010; Piatt et al., 2007).

46
47 Contamination of the marine environment comes from myriad sources, including influx
48 from rivers, agricultural and urban runoff, and long-range atmospheric transport (Amunsen et al.,
49 1992). Chemicals accumulate throughout marine waters (Lamborg et al., 2014; Lohmann et al.,
50 2007) and the species that inhabit them (Eagles-Smith et al., 2009; Finger et al., 2015). Some
51 elements are essential for metabolic and biological function, but in high concentrations are toxic
52 (e.g., copper (Cu); Lemos et al., 2013). However, exposure to low concentrations of non-essential
53 elements (e.g., mercury (Hg)) can contribute to poor health of individuals and populations. Elevated
54 trace element concentrations in birds has direct, and indirect health impacts, such as liver and
55 kidney damage (e.g. Cd; Hoffman et al., 2011; Nicholson and Osborn, 1983; Salamat et al., 2014),
56 reduced survival (Goutte et al., 2015), and changes in feather pigmentation (Pacyna et al., 2018).

57 While toxicity thresholds for some non-essential trace elements are poorly understood, especially in
58 birds, effect levels have been proposed for elements such as Hg (5 µg/g, though as high as 20 µg/g
59 in some marine species), lead (Pb; 4 µg/g), and cadmium (Cd; 0.1 – 5 µg/g) determined from
60 feathers samples of seabirds (Burger, 1993; Burger and Gochfeld, 2000b) however, these are highly
61 species-specific (Bond et al., 2015a). Analysis of muscle tissue from birds provides information on
62 trace element burden over a period of ~4 weeks (Lewis and Furness, 1991; Ramos and Gonzalez-
63 Solis, 2012), and is a reliable indicator of chemicals associated with metabolic function (Fromant et
64 al., 2016). Chemical pollutants in wildlife, including trace elements, are obtained from a variety of
65 sources, including prey (Bustamante et al., 1998), with some species experiencing heightened
66 exposure through the ingestion of plastic (Chua et al., 2014; Lavers et al., 2014; Sleight et al.,
67 2017). The nature of plastics to concentrate environmentally prevalent chemicals (Ashton et al.,
68 2010; Mato et al., 2001) and release them into an animal's tissues following ingestion (Tanaka et
69 al., 2013; Turner, 2018) is worrying as plastic ingestion has been documented in a variety of species
70 and across trophic levels, from invertebrate primary consumers (Dawson et al., 2018) to large
71 marine mammals (Fossi et al., 2018). The ingestion of plastic by wildlife has also been linked to
72 direct impacts, such as starvation (Pierce et al., 2004), as well as a range of lesser-known
73 physiological effects, including altered behaviour (e.g., lowered predator avoidance; Mattsson et al.,
74 2017; Seuront, 2018), tissue damage (Peda et al., 2016; Rochman et al., 2013) and changes in blood
75 chemistry (Lavers et al., 2019).

76

77 In seabirds, adults often provision chicks with prey collected from well-defined foraging
78 grounds adjacent to breeding colonies (Cleeland et al., 2014; Miller et al., 2018). Studies of Short-
79 tailed Shearwater (*Ardenna tenuirostris*) therefore provide information on the health of the Southern
80 Ocean, where parent birds forage during the chick rearing period (Cleeland et al., 2014), feeding on

81 low trophic level prey such as Antarctic krill (*Euphausia superba*; Skira et al., 1996). While
82 foraging for prey, adult birds mistakenly consume plastic and feed it to their chicks (Carey, 2011;
83 Waller et al., 2017), with >80% of Short-tailed Shearwater fledglings containing plastic debris
84 (Acampora et al., 2014; Cousin et al., 2015). It is unclear whether the chemicals detected in Short-
85 tailed Shearwater tissues (Honda et al., 1990; Lavers and Bond, 2013) originated primarily from
86 their prey, or ingested plastics as has been correlated in other seabird studies (Lavers and Bond,
87 2016a; Lavers et al., 2014).

88

89 In this study, we quantified plastic ingestion and trace element burdens of Short-tailed
90 Shearwater fledglings and compare with findings from similar data from 2011 (Lavers and Bond
91 (2013) to (a) examine the relationship between plastic ingestion, body condition, and trace element
92 burden in these birds, (b) provided updated data on the trace element concentration of Tasmanian
93 Sort-tailed Shearwaters, and (c) examine changes in plastic ingestion over time to perform a post
94 hoc power analysis of our current ability to assess these trends. We purchased 38 whole frozen
95 Short-tailed Shearwater fledglings from harvesters under recreational (individually held) collection
96 permits on Great Dog Island in the Furneaux Group, Tasmania, Australia (40.247°S, 148.249°E) in
97 late-April 2017. Birds were immediately frozen intact after collection and remained in frozen during
98 transportation and when stored at the Institute for Marine and Antarctic Studies aquatic animal
99 health laboratory. We recorded standard bird morphometric measurements (mass (± 10 g), wing
100 chord (unflattened and straightened; ± 1 mm), culmen (± 0.1 mm mm), head + bill (± 0.1 mm)) for
101 38 fledglings (approx. 80-90 days old) and followed the recommended protocol for quantifying
102 plastic in seabird necropsies (van Franeker, 2004). Ingested plastic items including micro-plastic
103 ($>1 \mu\text{m}$; Barnes et al., 2009) and visible macro-plastic ($> 5\text{mm}$ diameter; Provencher et al., 2017),
104 from the proventriculus and gizzard were categorised by type and colour following Provencher et

105 al. (2017), with a minor modification to the colour categories in order to be consistent with a long-
106 term monitoring study of Australian Procellariiformes (Lavers and Bond, 2016b; Puskic et al.,
107 2019). Plastic items were dried and weighed to the nearest 0.001 g using an electronic balance. The
108 whole breast muscle (pectoralis major; approximately 4.5 g (dry weight) in total) was removed and
109 stored individually in aluminium foil and sterile sealed bags at -20°C.

110

111 From a subset of fledglings (n = 20), the whole breast muscle sample was freeze-dried for
112 72 h to a constant weight, to determine dry matter and moisture content. Approximately 0.2000 g of
113 dried sample was homogenised to a fine powder using a mortar and pestle, placed in 100 ml glass
114 vials, and digested in 2 ml of concentrated nitric acid (HNO₃) (supplier details) with an antibumping
115 granule at 90°C for 2 h. The solution was analysed for 19 elements (Na, Mg, Al, S, Cr, Mn, Fe, Co,
116 Ni, Cu, Zn, As, Se, Rb, Sr, Mo, Ag, Cd and Pb) against the certified reference material dogfish
117 muscle (DORM-2; certified for Hg, and arsenic (As) from the National Research Council of
118 Canada) using a Sector Field Inductively Coupled Plasma Mass Spectrometer (ICP-MS Element 2,
119 Thermo Fisher, Bremen, Germany). International certifications for the remaining elements were
120 sourced from the Geological and Environmental Reference Materials database (GeoReM:
121 georem.mpch-mainz.gwdg.de; Jochum et al., 2005). To monitor analytical recovery and limit
122 erroneous data due to element rich matrices, a known quantity (10 ng/g) of a random sample digest
123 was spiked with a multi-element on top of the matrix. A total of 5 sample blanks were run for
124 quality control. The mean recovery across all elements in the spiked samples tested was ~80 -
125 130%, indicating the sample digest matrix had no detrimental effect on the results (Table S1).

126 Samples with a concentration below the detection limit of the ICP-MS were excluded
127 (136/380 total cases, 36%), and the remaining analytical duplicates averaged. Analysis of the
128 resulting dataset was completed using R (version 3.4.3; R Core Team, 2018) in RStudio (v.3.4.0,

129 Boston, Massachusetts, USA). The relationship between trace element concentrations and the mass
130 and number of ingested plastic items was examined using linear regression. We examined outliers
131 by calculating Cooks Distance, and as none had a value >3 , all data were retained (Cook, 1977,
132 1979). Relationships were considered significant when $p < 0.05$ and final results are reported as
133 $\mu\text{g/g}$ and expressed as mean \pm standard deviation (range). We extracted data from existing
134 publications on the frequency of occurrence, number and mass of plastic ingested by Short-tailed
135 Shearwaters ($n = 17$ papers). We separated these data by age class (adult and fledgling) and
136 excluded papers that did not report bird age (2 papers) or had sample sizes < 10 (5 papers; Table 1).
137 We conducted a power analysis (van Franeker and Meijboom, 2002) on this data following
138 Provencher et al. (2015), to estimate the sample size required to determine percent changes in the
139 frequency of occurrence, number of plastic pieces, and plastic mass in adult and fledgling samples
140 at 5, 10, 15, 25, 50, 75 and 100 %.

141
142 Plastic ingestion in Short-tailed Shearwaters appears to have remained consistent over time,
143 with a frequency of occurrence (FO) of $>80\%$ across most studies (Fig. 1, Table 1), though the
144 inconsistent monitoring of this species over time, age classes, and locations make this data difficult
145 to compare, and the collection methods (i.e., beach-washed, by-catch and road-killed birds) of each
146 of these studies vary and may influence our overall understanding of the interactions of these birds
147 and plastic (Rodríguez et al., 2018). Of the total fledglings necropsied ($n = 38$), 89.5% contained
148 plastic in their proventriculus (0.76 ± 1.34 items, range: 0-5; 19.5% of all plastic items) and/or
149 gizzard (3.24 ± 3.89 items, range: 0-13; 80.5% of all plastic items). Overall, the mean number of
150 plastic pieces recorded in individuals was 3.78 ± 3.79 items weighing 0.06 ± 0.06 g. The dominant
151 types of ingested plastic were hard fragments ($n = 103$ pieces, 75.7%) and pre-production pellets
152 ('nurdles'; 30 items, 22.1%), followed by thread, sheet and foam (1 item (0.7%) per type). The most

153 common colour of ingested plastic was white (87 items, 63.9%), followed by green (n = 24, 17.6%),
154 black (n = 20, 14.7%), blue (n = 3, 2.2%), red and yellow (n = 1, 0.7%) each (Table S2).. Short-
155 tailed Shearwater fledgling body mass varied greatly (583.5 ± 93.4 g; range 300-730 g), similar to
156 birds randomly collected as roadkill (522.64 ± 105.47 g; range 254-784 g; Rodríguez et al., 2018),
157 suggesting recreational harvesters selected birds at random. Fledgling morphometric measurements
158 were not related to plastic mass (bird mass: $F_{1,36} = 0.01$, $p = 0.91$, wing chord: $F_{1,36} = 2.22$, $p = 0.15$,
159 culmen: $F_{1,36} = 0.46$, $p = 0.50$, head + bill length: $F_{1,36} = 2.35$, $p = 0.13$,) or number of pieces (bird
160 mass: $F_{1,36} = 0.13$, $p = 0.73$, wing chord: $F_{1,36} = 1.22$, $p = 0.28$, culmen: $F_{1,36} = 1.68$, $p = 0.20$, head +
161 bill length: $F_{1,36} = 4.08$, $p = 0.05$). Plastic in high quantities can impact fledgling development in
162 some seabird species (Lavers et al., 2014), and has the capacity to introduce trace elements when
163 ingested (Browne et al., 2013; Lavers and Bond, 2016a; Luís et al., 2015). However, we found
164 minimal evidence of ingested plastic having a significant impact on Short-tailed Shearwater body
165 condition, in line with previous studies (Acampora et al., 2014; Carey, 2011; Cousin et al., 2015;
166 Rodríguez et al., 2018), and no evidence of a relationship between plastic and elevated trace
167 element concentrations in these birds. Short-tailed Shearwaters ingest relatively small amounts of
168 plastic compared to other Procellariiforms, such as Flesh-footed Shearwaters (*Ardenna carneipes*;
169 Lavers et al., 2014) and Laysan Albatross (*Phoebastria immutabilis*; Young et al., 2009), which
170 may explain our observation. Additionally, our ability to detect changes may be influenced by our
171 small sample size (n = 20 shearwaters analysed for trace elements).

172

173 We used data from 12 existing studies from 1969 – 2017 (Table 1) to conduct a power
174 analysis. Fledgling birds required lower sample sizes to detect changes in the frequency of
175 occurrence of plastic (to detect a 5% change in fledglings; n = 41, adults; n = 658). However, to
176 detect changes in the mass and number of plastic items ingested by these birds would require

177 sample sizes much larger than any study has reported to date (Table 2). Results from our power
178 analysis suggests changes within the frequency of occurrence of plastic ingested by Short-tailed
179 Shearwaters can be monitored with lower samples sizes in fledgling birds (5% change when $n = 41$)
180 compared to studies on adult birds (5% change when $n = 798$; Table 2). Fledgling seabirds are
181 regarded as reliable sentinels as they are provisioned by adults from well-defined foraging regions
182 and bird development is directly indicative of local conditions over a defined time-frame (Burger et
183 al., 2001; Provencher et al., 2019). In contrast, studies of adult birds can be used to explore the
184 biomagnification of contaminants over their life span. Mass and number of plastic pieces ingested
185 appears much harder to monitor (Table 2) due to the higher variability over time.

186

187 In a subset of Short-tailed Shearwater fledglings ($n = 20$), we detected measurable
188 concentrations of 10 trace elements in the dry breast muscle tissue (moisture content $62.5\% \pm 21.6$;
189 range: 8.25-92.94 %), the remaining nine elements were below the level of detection (LoD; Table
190 S1). Mean concentration of six non-essential trace elements (Cu, Zn, As, Rb, Sr and Cd), had
191 increased relative to the only comparable study on breast muscle samples of Tasmanian fledglings
192 in 2011 (Lavers and Bond, 2013; Table 3). A lack of long-term fledgling health or recruitment
193 monitoring and incidentally sparse datasets for trace element levels means we cannot draw trends or
194 predict what this data means for the overall health of these populations. Continued monitoring of
195 fledgling birds provides the best understanding of the anthropogenic impacts of the Southern Ocean,
196 where parent birds forage to feed young (Cleeland et al., 2014). As seabirds tend to bioaccumulate
197 trace elements over time monitoring of birds at a variety of life stages can also indicate the extent of
198 this pollution and what levels are frequent (Burger and Gochfeld, 2004; Mallory et al., 2010). As
199 the impacts of trace elements in seabirds appear to be highly species specific (Bond et al., 2015a),
200 and in some cases show little impact on individuals (Carravieri et al., 2018), future research on

201 Short-tailed Shearwaters should aim to determine effect levels of various trace elements. Many
202 trace elements, in high concentrations are associated with adverse health implications (Burger and
203 Gochfeld, 1985, 2000a). In Flesh-footed Shearwaters, high plastic ingestion is thought to be directly
204 related to mortality and the associated chemicals believed to incur morbidity (Fossi et al., 2018;
205 Lavers, 2015; Lavers et al., 2014). Trace element concentrations were unrelated to the mass or
206 number of plastic items ingested by fledgling shearwaters in our study and showed few significant
207 relationships with bird morphology (Table 4). Plastic production is expected to increase
208 (PlasticsEurope, 2018), therefore the number of species documented to be impacted by its ingestion
209 is also growing (Barnes et al., 2009; Gall and Thompson, 2015) and the indirect impacts of plastic
210 ingestion, such as the introduction of non-essential trace elements require urgent attention (Holmes,
211 2013).

212

213 The birds used in this study were destined for human consumption. Short-tailed Shearwaters
214 or Yolla are harvested in Tasmania annually and are of cultural significance to some aboriginal
215 Tasmanian communities (Hill et al., 1981). The harvesting of seabirds for their meat and fat is
216 documented in historical and archaeological records across the globe (Anderson, 1996; West and
217 Sim, 1995). While there is some concern regarding the consumption of high end consumers which
218 have high trace element concentrations (Bond et al., 2015b; Hightower and Moore, 2003; Johansen
219 et al., 2004; Johansen et al., 2006), globally, there is a lack of contaminant monitoring in these wild,
220 free-living and traditional food sources. In the Arctic, marine mammals destined for First Nations
221 consumption are assessed periodically (AFN, 2007; Moses et al., 2009). Dialogue between the
222 communities which consume unmonitored wildlife, the researchers who monitor these populations
223 and regulatory bodies to ensure traditional foods can be consumed safely while safeguarding
224 traditional practices and culture.

225

226 In the Northern hemisphere similar species such as the Northern Fulmer (*Fulmarus*
227 *glacialis*), are used in such a way that Ecological Quality Objectives (EcoQO) are put in place to
228 monitor plastic impacts in the local marine environment. EcoQO targets developed for Northern
229 Fulmars aim to see no more than 0.1 g of plastic in 10% of a population within a five year period
230 (OSPAR, 2008; van Franeker and Kühn, 2018). We lack similar international policies and targets
231 for pollution in Southern Hemisphere despite increasing data to develop these (Table 1). Due to the
232 number of studies on plastic ingestion in this species, the Short-tailed Shearwater can be presented
233 as one of the most studied Southern Ocean seabirds. The use of the Short-tailed Shearwater as
234 sentinel species may aid in the preservation of marine environments and conservation of cultural
235 heritage.

236

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251

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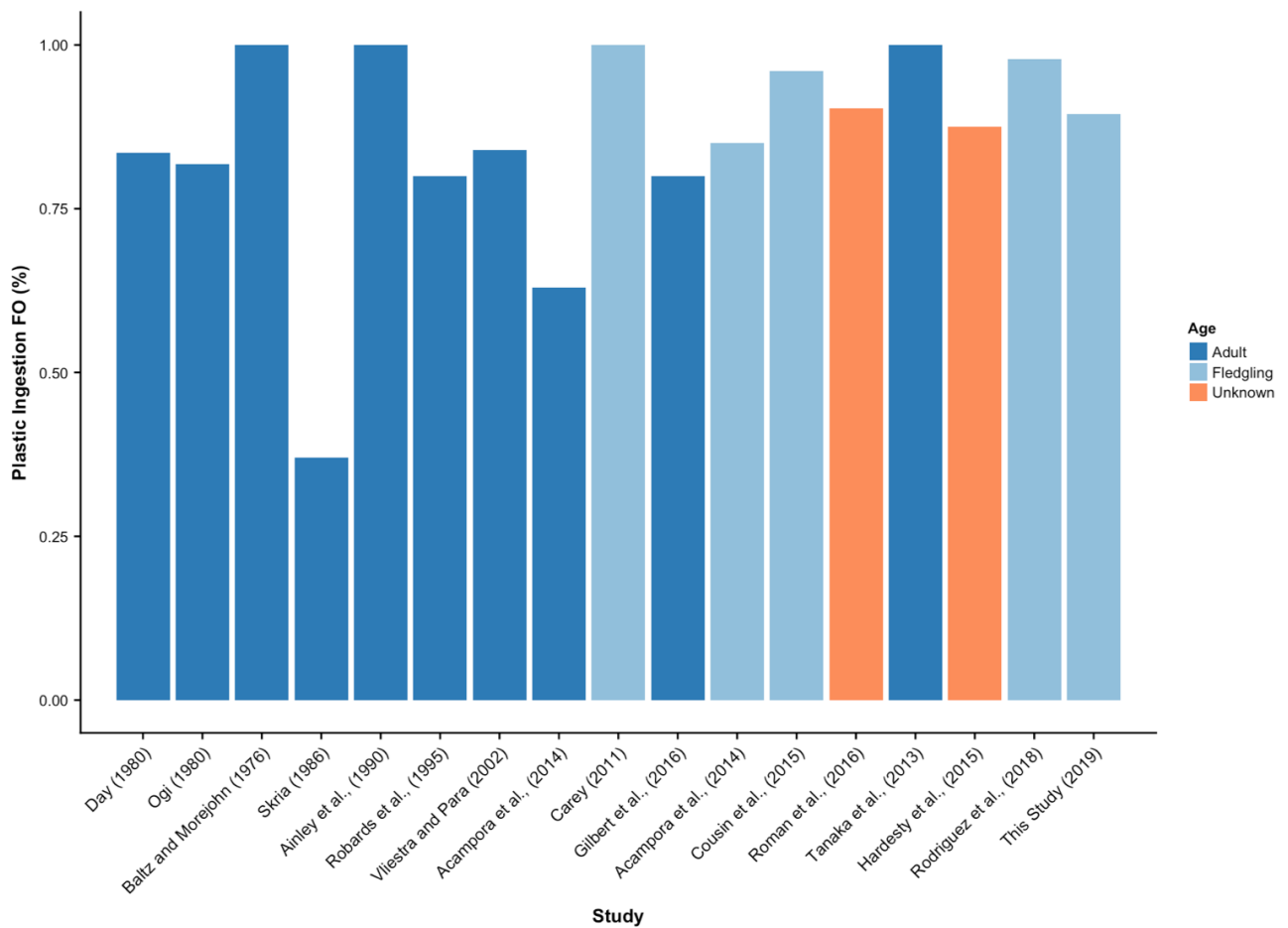
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515 **Fig. 1.** The frequency of occurrence (FO (%)) of plastic ingested by Short-tailed Shearwaters
 516 (*Ardenna tenuirostris*) over 30 years of global studies of birds at various life stages; adult (dark
 517 blue), fledgling (light blue) and unknown (orange) where studies did not define age. A total of 17
 518 studies are described, the details of each study are provided in Table 1. Only studies that provided
 519 FO% are included in this plot (n = 16 studies). Papers that provided data of both adults and
 520 fledglings were split (n = 1; Acampora et al., 2014).

521

522 **Table 1.** Review of published studies reporting plastic ingestion in Short-tailed Shearwaters (*Ardenna tenuirostris*). This data compliments that
 523 presented in Figure 1. FO: frequency of occurrence (%).

Year	Age	Location	n	FO (%)	Mean mass (g)	Mass range (g)	Mean pieces	Range pieces	Source
1969 - 1977	Adult	Sub-Antarctic, Pacific Ocean	200	83.5	0.10 ± 0.11	0 – 0.59	5.39 ± 5.72	0 – 32	(Day, 1980)
1970 - 1987	Adult	Sub-Antarctic, Pacific Ocean	324	81.8	0.14		8.79		(Ogi, 1990)
1974 - 1975	Adult	North Pacific Ocean	6	100					(Baltz and Morejohn, 1976)
1978	Adult	Tasmania, Australia	396	37			0-8		(Skira et al., 1996)
1984 - 1984	Adult	Pacific Ocean	1	100					(Ainley et al., 1990)
1988 - 1990	Adult	North Pacific Ocean	5	80					(Robards et al., 1995)
1997 - 2001	Adult	Bering Sea	330	84	0.114 ± 7.8		5.8 ± 0.4		(Vlietstra and Parga, 2002)
2003 - 2005	Adult	North Pacific Ocean	99		0.23 ± 0.18	0 – 0.9	15.1 ± 13.2		(Yamashita et al., 2011)
2010	Adult	Queensland, Australia	102	63			4.5		(Acampora et al., 2014)
2011	Fledgling	Victoria, Australia	67	100	0.113		7.6 ± 5.4		(Carey, 2011)
2011 - 2012	Adult	New South Wales, Australia	5	80		0.48 – 2.9		3 – 12	(Gilbert et al., 2016)
2012	Fledgling	Queensland, Australia	27	85			7.14		(Acampora et al., 2014)
2012	Fledgling	Tasmania, Australia	171	96	1.48		6.03 ± 0.3	0 – 30	(Cousin et al., 2015)
2013	Unknown	Eastern Australia	31	90.32			7.65		(Roman et al., 2016)
2013	Adult	North Pacific Ocean	12	100	0.24	0.04 – 0.59	27	1 – 9	(Tanaka et al., 2013)
2014	Unknown	Victoria, Australia	8	87.5			2.63	0 – 8	(Hardesty et al., 2015)
2015 - 2016	Fledgling	Victoria, Australia	140	97.85	0.15	0 – 1.04	6.46	0 – 26	(Rodríguez et al., 2018)
2017	Fledgling	Tasmania, Australia	38	89.5	0.06 ± 0.06	0 – 0.27	3.78 ± 3.79	0 – 15	This study

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526 **Table 2.** Results of a power analysis for ingested plastic by Short-tailed Shearwaters (*Ardenna tenuirostris*) showing the number of individuals
 527 required to detect changes in the frequency of occurrence (FO (%)), number and mass of ingested plastic.

Sample size required			
Age class & rate of change detected	FO (%)	Number of plastic pieces	Mass of plastic
Fledglings			
5%	41	530	2154
10%	11	145	5911
15%	5	71	2872
25%	<5	30	1221
50%	<5	11	440
75%	<5	7	266
100%	<5	5	195
Adults			
5%	798	5698	1476
10%	219	1563	405
15%	106	759	197
25%	45	323	84
50%	16	116	30
75%	10	70	18
100%	7	52	13

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532 **Table 3.** Review of trace element concentrations in the breast muscle of Short-tailed Shearwaters (*Ardenna tenuirostris*) reported as mean \pm standard
 533 deviation in $\mu\text{g/g}$ from juvenile birds harvested in Tasmania (Lavers and Bond, 2013; This Study) and adult birds collected from the North Pacific
 534 (Honda et al., 1990).
 535

Element	(Honda et al., 1990)	(Lavers and Bond, 2013)	This study
Ag	No Data	0.39 ± 0.20	<LoD
As	No Data	0.15 ± 0.14	1.04 ± 0.77
Br	No Data	1.27 ± 0.91	No Data
Cd	0.13 ± 0.07	0.01 ± 0.01	0.03 ± 0.03
Ce	No Data	0.02 ± 0.02	No Data
Co	No Data	0.01 ± 0.01	No Data
Cu	5.27 ± 0.64	4.00 ± 1.71	11.5 ± 2.69
Fe	63.1 ± 13.1	129.23 ± 38.69	No Data
Hg	0.04 ± 0.02	0.03 ± 0.00	No Data
Mg	No Data	326.54 ± 21.40	No Data
Mn	0.46 ± 0.11	0.53 ± 0.14	No Data
Pb	No Data	0.07 ± 0.13	No Data
Rb	No Data	1.39 ± 0.25	3.69 ± 0.45
Sb	No Data	<LoD (0.03)	No Data
Se	No Data	No Data	2.06 ± 0.31
Sr	No Data	0.13 ± 0.11	0.39 ± 0.47
Tl	No Data	<LoD (0.02)	No Data
V	No Data	<LoD (0.18)	No Data
Zn	16.9 ± 2.83	14.25 ± 6.11	62.31 ± 58.39

536

537 **Table 4.** P-values from linear models examining the relationships between elemental concentrations, morphometrics and plastic burden in Short-tailed
 538 Shearwaters.
 539

						540
	Bird Mass	Wing Chord Length	Culmen Length	Head + Bill Length	Plastic Number	Plastic Mass
As	0.35	< 0.001	0.68	0.61	0.27	0.50 ⁵⁴²
Cd	0.027	0.008	0.44	0.20	0.87	0.78 ⁵⁴³
Cu	0.11	0.03	0.30	0.009	0.49	0.024 ⁵⁴⁴
Mg	0.01	0.06	0.26	0.003	0.31	0.16 ⁵⁴⁵
Na	0.5	0.004	0.48	0.25	0.22	0.32 ⁵⁴⁶
Rb	0.48	0.73	0.57	0.98	0.73	0.98 ⁵⁴⁷
S	0.034	0.041	0.58	0.006	0.07	0.07 ⁵⁴⁸
Se	0.77	0.07	0.90	0.05	0.06	0.26 ⁵⁴⁹
Sr	0.66	0.40	0.59	0.97	0.48	0.19 ⁵⁵⁰
Zn	0.70	0.67	0.78	0.97	0.54	0.28 ⁵⁵¹